

Ecological site EX044B01C031 Limy Droughty (LyDr) 15-19" PZ Frigid North

Last updated: 3/03/2025
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General information

Provisional. A provisional ecological site description has undergone quality control and quality assurance review. It contains a working state and transition model and enough information to identify the ecological site.

MLRA notes

Major Land Resource Area (MLRA): 044B—Central Rocky Mountain Valleys

Major Land Resource Area (MLRA) 44B, Central Rocky Mountain Valleys, is nearly 3.7 million acres of southwest Montana. This MLRA borders two other MLRAs: 43B, Central Rocky Mountains and Foothills, and 46, Northern and Central Rocky Mountain Foothills.

The major watersheds of this MLRA are the Missouri and Yellowstone Rivers and their associated headwaters, such as the Beaverhead, Big Hole, Jefferson, Ruby, Madison, Gallatin, and Shields Rivers. Limited portions of the MLRA are west of the Continental Divide along the Clark Fork River. These waters allow for extensive irrigation for crop production in an area that is generally only compatible with rangeland and grazing. The Missouri River and its headwaters are behind several reservoirs used for irrigation water, hydroelectric power, and municipal water.

The primary land use of this MLRA is production agriculture (grazing, small grain production, and hay) with limited mining. Urban development is high, with large expanses of rangeland being converted to subdivisions for a rapidly growing population.

MLRA 44B consists of one Land Resource Unit (LRU) and 7 Climate-based LRU subsets. Annual precipitation ranges from a low of 9 inches to a high of near 24 inches. The driest areas tend to be in the valley bottoms of southwest Montana, in the rain shadow of the mountains. The wettest portions tend to be near the edges of the MLRA, where it borders MLRA 43B. Frost-free periods also vary greatly, with less than 30 days in the Big Hole Valley to approximately 110 days in the warm valleys along the Yellowstone and Missouri Rivers.

MLRA 44B's plant communities are highly variable but are dominated by a cool-season grass and shrub-steppe community on the rangeland and a mixed coniferous forest in the mountains. Warm-season grasses occupy an extremely limited extent and number of species in this MLRA. Most subspecies of big sagebrush are present, to some extent, across the MLRA.

LRU notes

LRU 01 Subset C Central Concept:

- Moisture Regime: Ustic
- Temperature Regime: Frigid
- Dominant Cover: rangeland (mixed grassland and sagebrush steppe)
- Representative Value (RV) of range of Effective Precipitation: 15 to 19 inches
- Representative Value (RV) of range of Frost Free Days: 75 to 105days

This LRU subset exists in northern portion of MLRA 44B particularly in Meagher, Powell, Broadwater, Lewis and Clark, Granite, and Deer Lodge Counties.

Classification relationships

Mueggler and Stewart. 1980. Grassland and Shrubland habitat types of Western Montana

1. *Stipa comata/Bouteloua gracilis* h.t.
2. *Agropyron spicatum/Bouteloua gracilis* h.t.

Montana Natural Heritage Program Vegetation Classification

1. *Stipa comata* - *Bouteloua gracilis* Herbaceous Vegetation

(STICOM – BOUGRA) Needle-and-thread/Blue grama

Natural Heritage Conservation Rank-G5 / S5

Edition / Author- 99-11-16 / S.V. Cooper,

EPA Ecoregions of Montana, Second Edition:

Level I: Northwestern Forested Mountains

Level II: Western Cordillera

Level III: Middle Rockies & Northern Great Plains

Level IV: Paradise Valley

Townsend Basin

Dry Intermontane Sagebrush Valleys

Shield-Smith Valleys

National Hierarchical Framework of Ecological Units:

Domain: Dry

Division: M330 – Temperate Steppe Division – Mountain Provinces

Province: M332 –Middle Rocky Mountain Steppe – Coniferous Forest – Alpine Meadow

Section: M332D – Belt Mountains Section

M332E – Beaverhead Mountains Section

Subsection: M332Ej – Southwest Montana Intermontane Basins and Valleys

M332Dk – Central Montana Broad Valleys

Ecological site concept

- Site does not receive any additional water
- Soils are
 - o Generally not saline or saline-sodic (limited extent)
 - o Moderately deep, deep, or very deep
 - o Typically less than 5 percent stone and boulder cover (15 percent maximum)
 - o Soil surface texture ranges from sandy loam to clay loam in surface mineral 4 inches
 - o Skeletal (greater than rock fragments) at a 10- to 20-inch soil control section
 - o Strongly or violently effervescent within surface mineral 4 inches; calcium carbonates will often increase with depth.
- Parent material is alluvium, slope alluvium, and colluvium (limited extent).

Associated sites

EX044B01C030	Limy (Ly) 15-19" PZ Frigid North The Limy ecological site occupies the same landscape position as the Limy Droughty.
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Similar sites

EX044B01C030	Limy (Ly) 15-19" PZ Frigid North The Limy ecological site differs by being not skeletal within 20 inches. The plant community is similar and shares a state and transition model.
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Table 1. Dominant plant species

Tree	Not specified
Shrub	(1) <i>Chrysothamnus viscidiflorus</i> (2) <i>Artemisia tridentata</i> ssp. <i>wyomingensis</i>

Herbaceous	(1) <i>Pseudoroegneria spicata</i> (2) <i>Hesperostipa comata</i>
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Legacy ID

R044BC031MT

Physiographic features

This ecological site occurs on slopes ranging from 0 to 60 percent; however, the representative slope is 4–10 percent. It is an area of dissected mountain valleys. The valleys are typically bordered by mountains trending from north to south. The parent material is tertiary valley fill and alluvium of mixed geology.

Table 2. Representative physiographic features

Landforms	(1) Intermontane basin > Terrace (2) Intermontane basin > Fan remnant (3) Intermontane basin > Alluvial fan
Ponding frequency	None
Elevation	4,500–6,000 ft
Slope	4–10%
Water table depth	42 in
Aspect	Aspect is not a significant factor

Climatic features

The Central Rocky Mountain Valleys MLRA has a continental climate. 50 to 60 percent of the annual long-term average total precipitation falls between May and August with the highest in May and June. Most of the precipitation in the winter is snow on frozen ground. Average precipitation for LRU 01 Subset C is 15 inches, and the frost-free period averages 75 to 105 days.

Table 3. Representative climatic features

Frost-free period (characteristic range)	75-105 days
Freeze-free period (characteristic range)	112-135 days
Precipitation total (characteristic range)	14-17 in
Frost-free period (actual range)	75-105 days
Freeze-free period (actual range)	112-135 days
Precipitation total (actual range)	13-20 in
Frost-free period (average)	90 days
Freeze-free period (average)	120 days
Precipitation total (average)	15 in

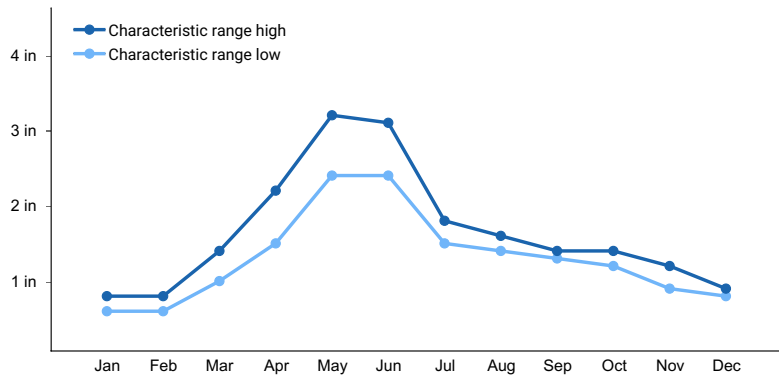


Figure 1. Monthly precipitation range

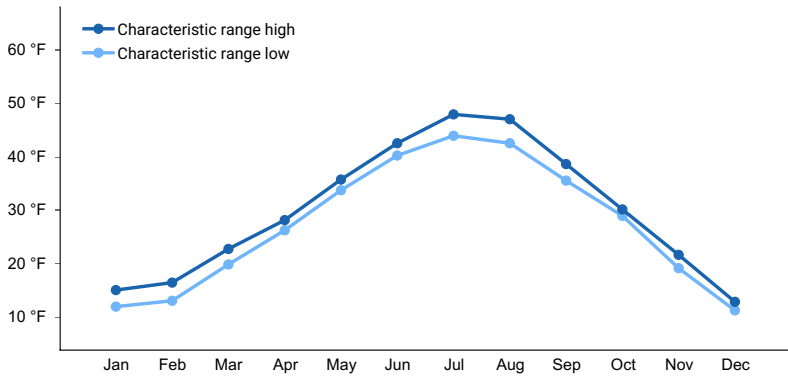


Figure 2. Monthly minimum temperature range

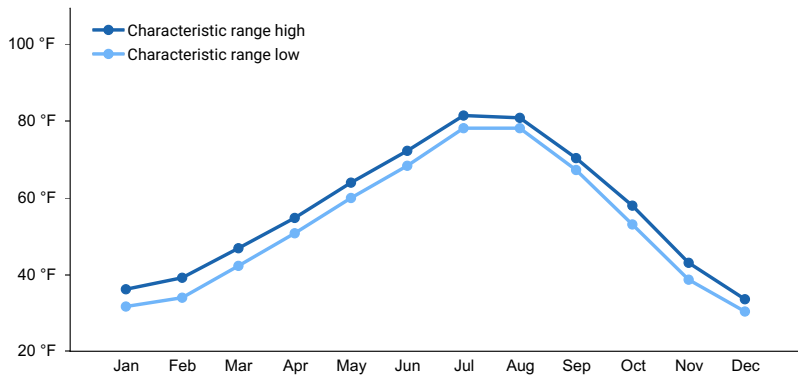


Figure 3. Monthly maximum temperature range

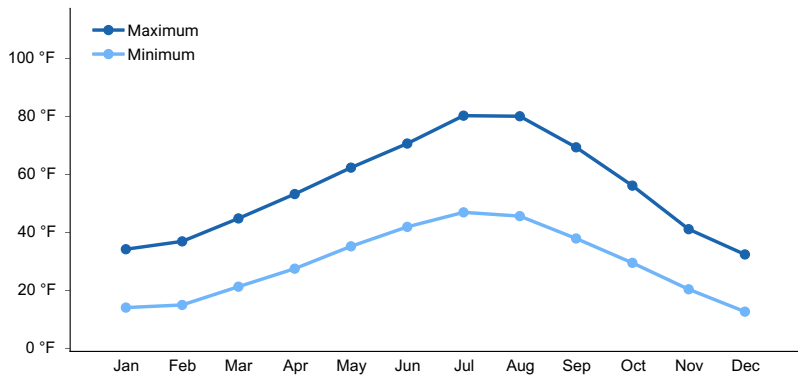


Figure 4. Monthly average minimum and maximum temperature

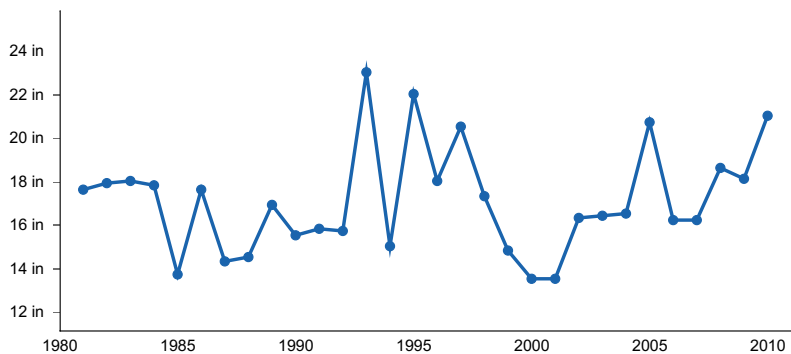


Figure 5. Annual precipitation pattern

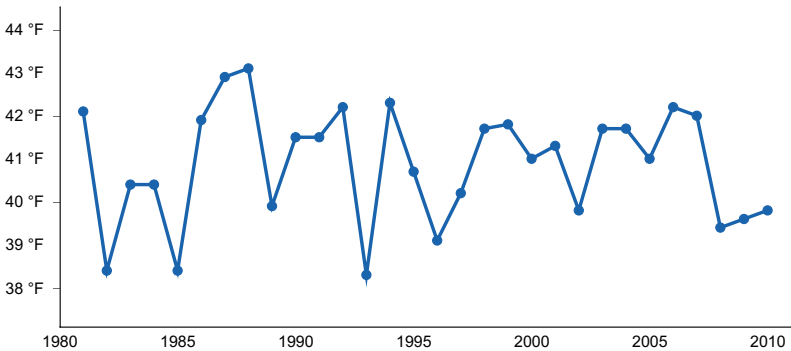


Figure 6. Annual average temperature pattern

Climate stations used

- (1) WILLSALL 8 ENE [USC00249023], Wilsall, MT
- (2) ANACONDA [USC00240199], Anaconda, MT
- (3) AUSTIN 1 W [USC00240375], Helena, MT
- (4) PHILIPSBURG RS [USC00246472], Philipsburg, MT
- (5) LENNEP 5 SW [USC00244954], White Sulphur Springs, MT
- (6) BOZEMAN MONTANA ST U [USC00241044], Bozeman, MT

Influencing water features

There are no water features influencing this site.

Wetland description

Site is not associate with wetland characteristics.

Soil features

These soils are moderately deep to very deep, moderately to moderately rapid permeability, and well drained. These soils are formed from alluvium, colluvium, and slope alluvium. The top four (4) inches of soil have strong to violent effervescence. The calcium carbonate (lime) concentration often increases with soil depth. The soil consists of loamy-skeletal material (which averages 35 percent or greater rock fragments by volume in the 10–20-inch layer). This skeletal material decreases the water-holding capacity of the site. Typically, soil surface textures consist of loam, sandy loam, and loamy sand textures. Soils are also typically gravelly, channery, or cobbly. Common soil series are Windham and Winspect. These soils may exist across multiple ecological sites due to natural variations in slope, texture, rock fragments, and pH. An onsite soil pit and the most current ecological site key are required to classify an ecological site.

Table 4. Representative soil features

Parent material	(1) Alluvium–limestone, sandstone, and shale (2) Colluvium–limestone, sandstone, and shale (3) Slope alluvium–limestone, sandstone, and shale
Surface texture	(1) Gravelly loam (2) Channery loam (3) Cobbly loam
Family particle size	(1) Loamy
Drainage class	Moderately well drained to well drained
Permeability class	Moderately slow to moderately rapid
Depth to restrictive layer	22–60 in
Soil depth	22–100 in
Surface fragment cover ≤3"	0–15%
Surface fragment cover >3"	5–10%
Available water capacity (0–40in)	2.2–5.2 in
Calcium carbonate equivalent (0–40in)	20–40%
Electrical conductivity (0–40in)	0–1 mmhos/cm
Sodium adsorption ratio (0–40in)	0–13
Soil reaction (1:1 water) (0–40in)	7.9–8.4
Subsurface fragment volume ≤3" (10–20in)	12–35%
Subsurface fragment volume >3" (10–20in)	10–20%

Table 5. Representative soil features (actual values)

Drainage class	Moderately well drained to well drained
Permeability class	Moderately slow to moderately rapid
Depth to restrictive layer	20–60 in
Soil depth	20–100 in
Surface fragment cover ≤3"	0–20%
Surface fragment cover >3"	0–15%
Available water capacity (0–40in)	2.2–5.8 in
Calcium carbonate equivalent (0–40in)	15–50%
Electrical conductivity (0–40in)	0–1 mmhos/cm
Sodium adsorption ratio (0–40in)	0–13
Soil reaction (1:1 water) (0–40in)	6.6–8.4
Subsurface fragment volume ≤3" (10–20in)	10–35%

Subsurface fragment volume >3" (10-20in)	15–35%
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Ecological dynamics

The reference plant community is dominated by bluebunch wheatgrass (*Pseudoroegneria spicata*) and needle and thread (*Hesperostipa comata*). Subdominant species trend toward rough fescue (*Festuca scabrella*), winterfat (*Krascheninnikovia lanata*), and Indian ricegrass (*Achnatherum hymenoides*). This potential is suggested by investigations showing a predominance of perennial grasses on near-pristine range sites (Ross et al., 1973).

The Limy Droughty ecological site LRU 01 Subset B occurs across a relatively small landscape, though slight variations exist due to elevation, frost-free days, and relative effective annual precipitation. Bluebunch wheatgrass, for example, occupies all known combinations of elevation and climate in this subset. Conversely, colder, wetter sites within this subset often exhibit slight increases in rough fescue and Wyoming big sagebrush. Unique to the Limy Droughty and Limy ecological sites of Subset C, there is an increased probability of Idaho fescue as Relative Effective Annual Precipitation increases to beyond 17 inches. This seems to be the threshold for precipitation to push the concentration of carbonates deeper into the soil profile. This correlation is also expressed in higher amounts of rough fescue.

A shift to the dominance of shrubs may occur in response to improper grazing management, drought, or where Wyoming big sagebrush occurs due to a lack of fire. Shrub encroachment by a variety of species, including broom snakeweed (*Gutierrezia sarothrae*), prairie sagewort (*Artemisia frigida*), Wyoming big sagebrush (*Artemisia tridentata* ssp. *wyomingensis*), rubber rabbitbrush (*Ericameria nauseosa*), yellow rabbitbrush (*Chrysothamnus viscidiflorus*), and plains prickly pear (*Opuntia polyacantha*) occurs within this site as the mid-stature bunchgrasses decrease. Shrub dominance and grass loss are associated with soil erosion and, ultimately, thinning of the native soil surface. Subsequent loss of soil could lead to a Degraded State. All states could also lead to the Invaded State when there is a lack of weed prevention and control measures.

Historical records indicate that prior to the introduction of livestock (cattle and sheep) during the late 1800s, elk and bison grazed this ecological site. Due to bison's nomadic nature and herd structure, grazed areas received periodic high intensity, short duration grazing pressure. The gold boom in the 1860s brought the first herds of livestock overland from Texas, and homesteaders began settling the area. During this time, cattle were the primary domestic grazers in the area. In the 1890s, Montana sheep production began to increase (by more than 400 percent) and dominated the livestock industry until the 1930s. Since the 1930s, cattle production has dominated the livestock industry in the region (Wyckoff and Hansen 2001).

Natural fire as well as prescribed fire utilized by indigenous peoples was a major ecological driver of this not only this ecological site but the entire MLRA. Indigenous peoples have utilized fire on this ecological site for thousands of years prior to European settlement as a means to move wildlife populations for harvest (Roos Christopher I. et al. 2018). Fire tended to restrict tree and shrub growth to small patches and promoted an herbaceous plant community. The natural fire return interval was highly variable, but it was likely shorter than 30 years. With the historically recent, since 1910, suppression of fire, shrubs and coniferous trees have increased significantly.

Due to the slightly alkaline to moderately alkaline nature and rockiness of the soils on this site, the potential for crop production is extremely limited. This ecological site remains mostly unconverted.

Some of the major invasive species that can occur on this site include spotted knapweed (*Centaurea stoebe*), leafy spurge (*Euphorbia esula*), and cheatgrass (*Bromus tectorum*). Nonnative invasive weeds are generally not common in most of this ecological site and tend to occupy limited areas in small patches near traditional watering facilities, along roads, and other areas that receive high soil disturbance.

Plant Communities and Transitions

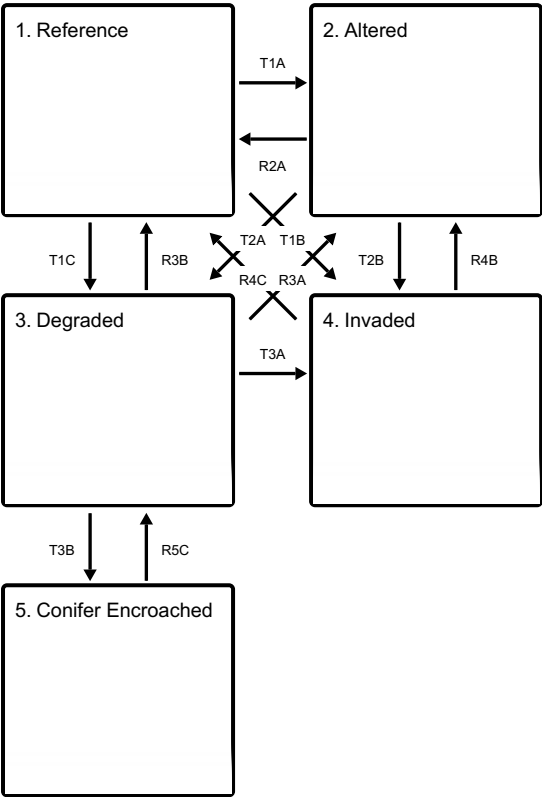
A state and transition model (STM) for this ecological site is depicted below. Thorough descriptions of each state, transition, plant community, and pathway follow the model. This model is based on available experimental research, field data, field observations, and interpretations by experts. It is likely to change as knowledge increases.

The plant communities within the same ecological site will differ across the MLRA due to the naturally occurring variability in weather, soils, and aspect. The biological processes on this site are complex; therefore, representative

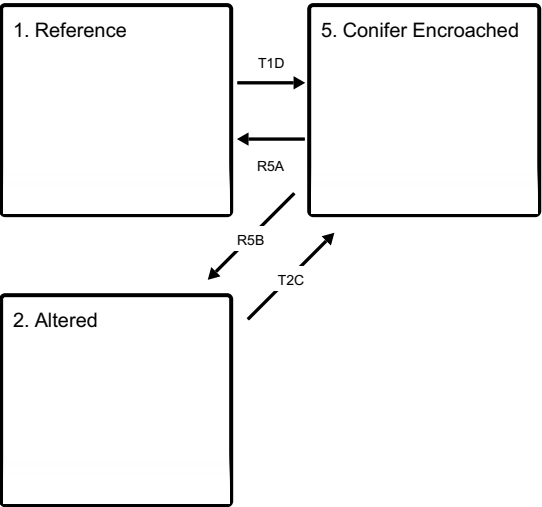
values are presented in a land management context. The species lists are representative and are not botanical descriptions of all species occurring, or potentially occurring, on this site. They are intended to cover the core species and the known range of conditions and responses.

State and transition model

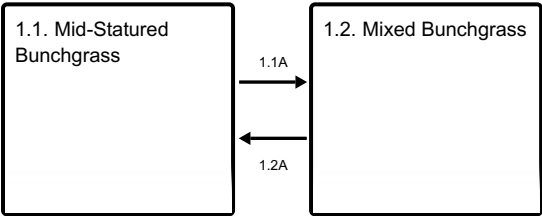
Ecosystem states



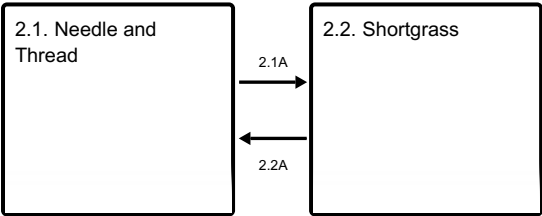
States 1, 5 and 2 (additional transitions)



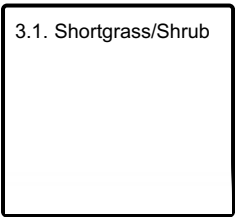
State 1 submodel, plant communities



State 2 submodel, plant communities



State 3 submodel, plant communities



State 4 submodel, plant communities

4.1. Invaded

State 5 submodel, plant communities

5.1. Conifer
Encroached

State 1 Reference

The Reference State of this ecological site consists of two (2) known potential plant communities: the Mid-Statured Bunchgrass Community and the Mixed Bunchgrass Community. These are described below but are generally characterized by a mid-statured, cool-season grass community with limited shrub production. Community 1.1 is dominated by bluebunch wheatgrass and is considered the reference, while Community 1.2 has a codominance of bluebunch and needle and thread with an increase in yellow rabbitbrush and Wyoming big sagebrush. These communities may meld into each other due to the varying conditions that occur in Southwest Montana, particularly during dry cycles where the needle and thread growth cycle takes better advantage of the limited moisture.

Community 1.1 Mid-Statured Bunchgrass

Bluebunch wheatgrass (*Pseudoroegneria spicata*) and needle and thread (*Hesperostipa comata*) are typically dominant. Rough fescue may also present in this community as a subordinate with Indian ricegrass (*Achnatherum hymenoides*) and winterfat (*Krascheninnikovia lanata*). Shrub species (big sagebrush, prairie sagewort, and broom snakeweed) remain a minor part of the community. In areas where the soil texture is coarser, spineless horsebrush (*Tetradymia canescens*) may occupy a small niche. Sandberg bluegrass (*Poa secunda*), prairie Junegrass, and dryland sedges are also common. This plant community occurs in areas that receive properly managed livestock or wildlife grazing. Bluebunch wheatgrass lacks resistance to grazing during the critical growing season (spring) and will decline in vigor and production if grazed in the critical growing season more than one year in three (Wilson et al. 1960). The Mid-Statured Bunchgrass Community is moderately resilient and will return to dynamic equilibrium following a relatively short period of stress (such as drought or short-term improper grazing); provided a return to favorable or normal growing conditions and properly managed grazing. As discussed in the Ecological Dynamics section, the natural fire regime restricted shrubs to relatively small portions of the Mid-Statured Bunchgrass Community. Shrub species present may include Wyoming big sagebrush, spineless horsebrush, winterfat, tarragon (*Artemisia drucunculus*), and prairie sagewort. Infrequent fire probably maintained big sagebrush communities as open, seral stands of productive herbaceous species with patches of big sagebrush.

Dominant plant species

- Wyoming big sagebrush (*Artemisia tridentata* ssp. *wyomingensis*), shrub
- yellow rabbitbrush (*Chrysothamnus viscidiflorus*), shrub
- winterfat (*Krascheninnikovia lanata*), shrub
- bluebunch wheatgrass (*Pseudoroegneria spicata*), grass
- green needlegrass (*Nassella viridula*), grass
- needle and thread (*Hesperostipa comata*), grass
- hairy false goldenaster (*Heterotheca villosa*), other herbaceous
- American vetch (*Vicia americana*), other herbaceous
- dotted blazing star (*Liatris punctata*), other herbaceous
- lupine (*Lupinus*), other herbaceous

Table 6. Annual production by plant type

Plant Type	Low (Lb/Acre)	Representative Value (Lb/Acre)	High (Lb/Acre)
Grass/Grasslike	760	980	1190
Forb	60	100	140
Shrub/Vine	80	120	140
Total	900	1200	1470

Table 7. Ground cover

Tree foliar cover	0%
Shrub/vine/liana foliar cover	10-20%
Grass/grasslike foliar cover	55-75%
Forb foliar cover	5-15%
Non-vascular plants	0-1%
Biological crusts	0-1%
Litter	20-30%
Surface fragments >0.25" and <=3"	0-15%
Surface fragments >3"	0-10%
Bedrock	0%
Water	0%
Bare ground	7-15%

Table 8. Soil surface cover

Tree basal cover	0%
Shrub/vine/liana basal cover	1-5%
Grass/grasslike basal cover	10-15%
Forb basal cover	1-3%
Non-vascular plants	0-1%
Biological crusts	0-1%
Litter	20-30%
Surface fragments >0.25" and <=3"	0-15%
Surface fragments >3"	0-10%
Bedrock	0%
Water	0%
Bare ground	7-15%

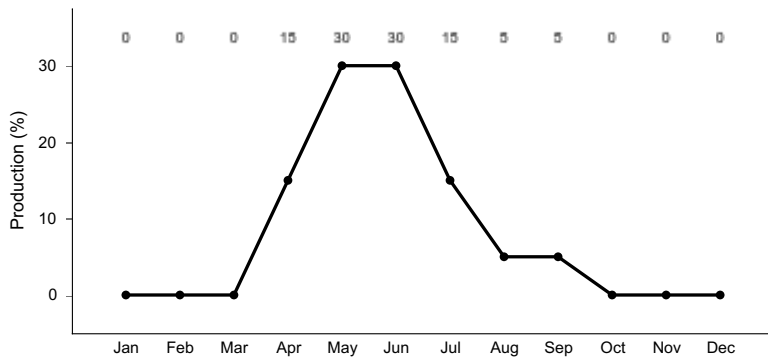


Figure 8. Plant community growth curve (percent production by month). MT44B032, Dry Uplands. Cool season grass dominated system. Most dry, upland sites located within MLRA 44B LRU A are characterized by early season growth which is mostly complete by Mid-July. Limited fall "green-up" if conditions allow..

Community 1.2 Mixed Bunchgrass

Western wheatgrass, thickspike wheatgrass, and needle and thread tolerate grazing pressure better than bluebunch wheatgrass and rough fescue. The growing point for bluebunch wheatgrass and rough fescue is several inches above the ground, making it very susceptible to continued close grazing, while the growing points for western and thickspike wheatgrass and needle and thread tend to be near the plant base (Smoliack et al., 2006). These plants increase in composition when less palatable and less grazing-tolerant plants decrease due to improper grazing management. Needle and thread, western/thickspike wheatgrass, and bluebunch wheatgrass share dominance in the Mixed Bunchgrass Community. Other grass species that are more tolerant of grazing and are likely to increase in number compared to the Reference Plant Community include Sandberg bluegrass (*Poa secunda*), prairie Junegrass, thickspike wheatgrass (*Elymus lanceolatus*), and blue grama (*Bouteloua gracilis*). Increaser forbs include Western yarrow, spiny phlox (*Phlox hoodii*), scarlet globemallow (*Sphaeralcea coccinea*), hairy false goldenaster (*Heterotheca villosa*), and pussytoes (*Antennaria* spp.). Prairie sagewort is a shrub that also increases under prolonged drought or heavy grazing and can respond to precipitation that falls in July and August. Heavy, continuous grazing will reduce plant cover, litter, and mulch. The timing of grazing is important on this site because of the moisture limitations beyond June, especially on the drier sites. Bare ground will increase, exposing the soil to erosion. Litter and mulch will be also reduced as plant cover declines. The presence of needle and thread and western wheatgrass will increase. If the dry weight of bluebunch wheatgrass is reduced to less than 30 percent, it will be challenging for the site to recover to the Mid-Statured Bunchgrass Community. When canopy cover decreases, the risk of soil erosion increases. Deteriorating soil conditions lead to a loss of organic matter, reduced litter, and decreased soil fertility. The degraded soil conditions make it harder to reestablish bluebunch wheatgrass and return to the Mid-Statured Bunchgrass Community. The Mixed Bunchgrass Community is an at-risk plant community for the Reference State. When overgrazing continues, increaser species such as needle and thread and native forb species will become more dominant, and this triggers the change to the Altered State or the Degraded State. This community can be managed toward the Mid-Statured Bunchgrass Community using prescribed grazing and strategic weed control, if present. It will take several years to achieve this recovery based on growing conditions, the vigor of remnant bluebunch wheatgrass plants, and the aggressiveness of the weed treatments.

Pathway 1.1A Community 1.1 to 1.2

Bluebunch wheatgrass loses vigor with improper grazing or extended drought. When vigor declines enough for plants to die or become smaller, species with higher grazing tolerance increase in vigor and production as they access the resources previously used by bluebunch wheatgrass. The decrease in species composition by weight of Mid-Statured Bunchgrasses to less than 50 percent indicates that the plant community has shifted to the Mixed Bunchgrass Community (1.2). The driver for community shift 1.1A is improper grazing management or prolonged drought. This shift is triggered by the loss of vigor of bluebunch wheatgrass, soil erosion, or prolonged drought coupled with improper grazing. Blaisdell (1958) stated that drought and warmer-than-normal temperatures are known to advance plant phenology by as much as one month. During drought years, plants may be especially sensitive or reach a critical stage of development earlier than expected. Since needle and thread normally heads out in June and bluebunch wheatgrass in July, this should be taken into consideration when planning grazing

management.

Pathway 1.2A

Community 1.2 to 1.1

The Mixed Bunchgrass Community (1.2) will return to the Mid-Statured Bunchgrass Community (1.1) with proper grazing management and appropriate grazing intensity. Favorable moisture conditions will facilitate or accelerate this transition. It may take several years of favorable conditions for the community to transition back to a bluebunch dominated state. The driver for this community shift (1.2A) is the increased vigor of bluebunch wheatgrass, to the point that it represents more than 50 percent of species composition. The trigger for this shift is the change in grazing management favoring bluebunch wheatgrass. This includes conservative grazing management styles such as deferred or rest rotations utilizing moderate grazing (less than 50 percent use) combined with favorable growing conditions such as cool, wet springs. These systems promote increased soil organic matter, which promotes microfauna and can increase infiltration rates. Conversely, long periods of rest at a time when this state is considered stable may not result in an increase in bluebunch wheatgrass. These long periods of rest or underutilization may actually drive the system to a lower level of stability by creating large amounts of standing biomass, dead plant caudex centers, and gaps in the plant canopy (Noy-Meir 1975).

State 2

Altered

This state is characterized by having less than 20 percent bluebunch wheatgrass by dry weight. It is represented by two (2) communities that differ in the percent composition of needle and thread, production, and soil degradation. Production in this state can be similar to that in the Reference State (1). Some native plants tend to increase under prolonged drought and/or heavy grazing practices. A few of these species may include needle and thread, Sandberg bluegrass, scarlet globemallow, hairy false goldenaster, and prairie sagewort.

Community 2.1

Needle and Thread

Long-term grazing mismanagement with continuous growing-season pressure will reduce the total productivity of the site and lead to an increase in bare ground. Once plant cover is reduced, the site is more susceptible to erosion and degradation of soil properties. Soil erosion or reduced soil health will result in reduced plant production. This soil erosion or loss of soil fertility indicates the transition to the Altered State because it creates a threshold requiring energy input to return to the Reference State. Transition to the Needle and Thread Community may be exacerbated by extended drought conditions. Needle and thread dominates this Community. Bluebunch wheatgrass makes up less than 20 percent of the species composition by dry weight, and the remaining bluebunch wheatgrass plants tend to be scattered and low in vigor. Invasive species will become more common, increasing competition for bluebunch wheatgrass. This makes it difficult for bluebunch wheatgrass to quickly respond to a change in grazing management alone. Therefore, an input of energy is required for the community to return to the Reference State. Wind and water erosion may be eroding soil from the plant interspaces. Soil fertility is reduced, and soil surface erosion resistance has declined compared to the Reference State. This community crossed a threshold compared to the Mixed Bunchgrass Community due to the erosion of soil, vegetation composition, loss of soil fertility, or degradation of soil conditions. This results in a critical shift in the ecology of the site. The effects of soil erosion can alter the hydrology, soil chemistry, soil microorganisms, and soil structure to the point where intensive restoration is required to restore the site to another state or community. Changing grazing management alone cannot create sufficient improvement to restore the site within a reasonable time frame. With decreased grazing pressure, a needle and thread plant community did not change species composition, but the content of the soil carbon increased (Dormaer 1997). It will require a considerable input of energy to move the site back to the Reference State. This state has lost soil or vegetation attributes to the point that recovery to the Reference State will require reclamation efforts, i.e., soil rebuilding, intensive mechanical treatments, and/or reseeding. The transition to this state could result from overgrazing, especially repeated early-season grazing coupled with extensive drought. If heavy grazing continues, plant cover, litter, and mulch will continue to decrease, and bare ground will increase, exposing the soil to accelerated erosion. Litter and mulch will move off-site as plant cover declines. The Needle and Thread Community will then shift to the Shortgrass Community. Continued improper grazing will drive the community to a Degraded State. Introduction or expansion of invasive species will further drive the plant community into the Invaded State. With increased bare ground associated with less vigorous plants, the Needle and Thread Community is susceptible

to invasion from native coniferous trees and tall shrubs, leading to the Conifer Encroached State.

Community 2.2

Shortgrass

With continued mismanagement of grazing, especially coupled with prolonged drought, needle and thread will decrease in vigor, causing a shift to the more degraded shortgrass community. The bunchgrasses will decline in production as plants die or become smaller, and species with higher grazing tolerance (such as thickspike and western wheatgrass) will increase in vigor and production as they respond to resources previously used by the bunchgrasses. These less desirable, shorter-rooted species will become co-dominant with the taller bunchgrasses. Shrubs will become more competitive for limited moisture as bare ground and soil erosion increase. Bluebunch wheatgrass is reduced to less than 10 percent of the plant community's production. Rough fescue is completely removed from the community. Continued degradation will result in a transition to the Degraded State (3), Invaded State (4), or Conifer Encroached State.

Pathway 2.1A

Community 2.1 to 2.2

The driver for community shift 2.1A is continued improper grazing management. This shift is triggered by the continued loss of bunchgrass vigor, especially since needle and thread is the remaining mid-statured bunchgrass. The short-statured grasses will become more competitive and will become co-dominant with the bunchgrasses. Shrubs will increase their canopy cover. Bluebunch is reduced to less than 10 percent production, and rough fescue is absent.

Pathway 2.2A

Community 2.2 to 2.1

If proper grazing management is implemented, needle and thread and bluebunch wheatgrass may regain their vigor and move toward the Needle and Thread Community. This will give grasses an advantage over invading shrubs before too much competition takes place. The advantage to grasses comes from following a conservative grazing plan where utilization is reduced and rest or deferment is incorporated since the transition from Plant Community 2.1 to Plant Community 2.2 is likely caused by repeated heavy utilization. Forage production can increase by an average of 35 percent on western ranges when converting heavy to moderate utilization (Van Poolen and Lacey 1979). Shrub removal and favorable growing conditions can accelerate this process. If the site contains Wyoming big sagebrush, low-intensity fire or mechanical treatment could reduce shrub competition and allow for increased vigor and the reestablishment of grass species (Wambolt 1986).

State 3

Degraded

The Degraded State is described by a single plant community consisting of nearly equal components of increaser grasses, shrubs, and forbs. Large patches of bare ground exist, with areas of erosional pedestalling and terracettes common. Lesser spikemoss (*Selaginella densa*) exists between plant bases as a reaction to the increased bare ground.

Community 3.1

Shortgrass/Shrub

Soil loss continues or increases to the point that native perennial grasses make up less than half of the annual dry weight production. Grass and forb cover may be very sparse or clumped. Weeds, annual species, cacti, or shrubs dominate the plant community. Mid-stature perennial bunchgrass species may exist, but only in small patches. This occurs as a result of overgrazing due to a failure to adjust stocking rates to declining forage production. In the most severe stages of degradation, there is a significant amount of bare ground, and large gaps occur between plants. Large patches of prickly pear cactus are common. Potential exists for soils to erode to the point that irreversible damage may occur. This is a critical shift in the ecology of the site. Soil erosion combined with a lack of organic matter deposition due to sparse vegetation creates changes to the hydrology, soil chemistry, soil microorganisms, and soil structure to the point where intensive restoration is required to restore the site to another state or

community. Changing management cannot create sufficient change to restore the site within a reasonable time frame. This state is characterized by soil surface degradation and little plant soil surface cover. Shrub canopy cover is usually greater than 25 percent. In this plant community, big sagebrush is replaced with a dominant community of broom snakeweed, rabbitbrush, prairie sagewort, and plains prickly pear cactus. This state has lost soil and vegetation attributes to the point that recovery to the Reference State will require reclamation efforts, i.e., soil rebuilding, intensive mechanical treatments, and/or reseeding. This plant community may be in a terminal state and will not return to the reference state because of degraded soil conditions and the loss of higher successional native plant species. Significant soil reconstruction and reseeding will be required to restore the plant community to its Reference State. Key factors in the approach to transition include a decrease in grass canopy cover and production, an increase in shrub canopy cover, increases in mean bare patch size, increases in soil crusting, decreases in cover of cryptobiotic crusts, decreases in soil aggregate stability, and/or evidence of erosion, including water flow patterns and litter movement.

State 4 Invaded

The Invaded State is identified as being in the exponential growth phase of invader abundance, where control is a priority. Dominance (or relative dominance) of noxious or invasive species reduces species diversity, forage production, wildlife habitat, and site protection. A level of 10 percent invasive species composition by dry weight indicates that a substantial energy input will be required to create a shift to the grassland state (herbicide, mechanical treatment), even with a return to proper grazing management or favorable growing conditions. In some instances, carefully targeted grazing (sometimes in combination with other treatments) can reduce or maintain the composition of invasive species. If the invasive or noxious species continue to thrive without mechanical, biological, or chemical control methods, this may become a terminal state. Once the weed reaches its maximum population level for this site, effective control is unlikely without massive resource inputs. After invading species have established and spread, the site's ecological processes (as defined in the Interpreting Indicators of Rangeland Health) may change (Walker and Smith 1997). The lack of native plants reduces site stability and hydrologic function.

Community 4.1 Invaded

Communities in this state may be structurally indistinguishable from the bunchgrass state except that invasive/noxious species exceed 10 percent of species composition by dry weight. This state may also include a community similar to the Degraded State except that invasive/noxious species exceed 20 percent of species composition by dry weight. Although there is no research to document the level of 20 percent, this is estimated to be the point in the invasion process following the lag phase based on interpretation of Masters and Sheley 2001.

Early in the invasion process there is a lag phase where the invasive plant populations remain small and localized for long periods before expanding exponentially (Hobbs and Humphries 1995). Production in the invaded community may vary greatly. A site dominated by Kentucky bluegrass or spotted knapweed, where soil fertility and chemistry remain near reference, may have production near that of the reference community. A site with degraded soils and an infestation of cheatgrass may produce only 10 to 20 percent of the reference community. Once invasive species dominate the site, either in species composition by weight or in their impact on the community the threshold has been crossed to the Invaded State. As invasive species such as spotted knapweed, cheatgrass, and leafy spurge become established, they become very difficult to eradicate. Therefore considerable effort should be placed in preventing plant communities from crossing a threshold to the Invaded State through early detection and proper management. Preventing new invasions is by far the most cost-effective control strategy, and typically places an emphasis on education. Control measures used on the noxious plant species impacting this ecological site include chemical, biological, and cultural control methods. The best success has been found with an integrated pest management (IPM) strategy that incorporates one or several of these options along with education and prevention efforts (DiTomaso 2000).

State 5 Conifer Encroached

The Conifer Encroached State may contain as many as 4 different phases. The Early Phase, Mid Phase, Late Phase, and Closed Phase are defined by the amount of encroachment and age class of the stand. This state

typically occurs in response to a combination of long-term fire suppression, grazing history, and increase atmospheric carbon dioxide. The trigger for transition is a coniferous expansion of greater than 1 stem per hectare.

Community 5.1

Conifer Encroached

Rocky Mountain juniper (*Juniperus scopulorum*), ponderosa pine (*Pinus ponderosa*), and Douglas fir (*Pseudotsuga menziesii*) encroachment is common on this ecological site. The encroachment is generally focused in areas where the mountains of MLRA 44B transition quickly to MLRA 43B. Conifers are absent on this ecological site, but, if present, will not exceed one (1) stem per hectare. All states in this ecological site can transition to the Conifer Encroached State. Encroachment is most likely to occur in the Altered State, where there is an increase in bare ground due to a combination of factors that allows seed-to-soil contact with reduced competition. Fire suppression and improper grazing management are the two most common triggers. The exact mode in which conifers begin to encroach varies; however, the trend points to a combination of 1 or more of the following: repeated moderately heavy to heavy grazing; reduced (non-existent) fire frequency; increased atmospheric carbon; and a generally warmer climate compared to that of pre-settlement. When heavy grazing occurs, areas in the plant canopy open, allowing for seed dispersal by bird or overland flow via rills on neighboring sites. The effects of conifer encroachment are not immediately noticeable, but over time, as the conifer canopy increases, light and water interception increase, which reduces opportunities for herbaceous plants. One paper suggests that for precipitation to penetrate the juniper canopy, events must be greater than 0.30 inches (Barrett, 2007). Increased tree canopy creates perching sites for predators, which reduces site suitability for greater sage grouse. Studies based in an area similar to the Rocky Mountain juniper community of Montana suggest following a phased approach to characterizing the juniper stand (Miller et al., 2000). Not unlike the western juniper community discussed in Miller et al., the Conifer Encroached Communities of Montana exhibit 3 or 4 different phases based, at this time, on qualitative experience. Phase I (Early) is defined by actively expanding conifer cover with generally less than 10 percent canopy cover and the trees' limbs generally touching the ground. This early stage generally has not completely lost its hydrologic functions, but herbaceous plant communities may show signs of reduced production and species richness. Control methods include mechanical removal and prescribed fire. Prescribed fire is effective in this phase as this phase still contains the necessary native plants for recovery. Phase II (Midphase) is still actively expanding, but canopy cover may reach 15 to 25 percent, and due to the more mature trees, seed production is very high. This Midphase begins to highly restrict herbaceous and shrubby plants, and junipers tend to be codominant. Hydrology has departed from the Reference State, with rills becoming longer and, in isolated areas, erosional gullies possible. Control methods for the Midphase should focus on mechanical treatment, as there is a high risk of catastrophic and potentially sterilizing fire. Phase III (Late Phase) is where conifer cover exceeds 25 percent and has slowed as a forest condition. Lower tree limbs will begin to die, and the shrub cover is nearly gone. Traveling through this community is increasingly difficult. Conifers become the dominant plant, with herbaceous plant production greatly decreasing. Bare ground has increased and hydrologic function is nearly lost compared to a grass or shrub community. The late phase should focus more on restoration than control, as the necessary plants will likely not be present to cross the threshold back to a rangeland situation. Soil stability and hydrologic function are lacking in this phase, mechanical juniper removal will be required. Phase IV (Closed Phase) is the steady state forest, where the system is nearly devoid of rangeland plants. The trees stop producing seed and begin to close in on each other. The closed phase is extremely rare in this LRU. This phase takes upwards of 100 years to occur and even under suppression, fire will control these sites. Management often occurs before trees are allowed to reach this phase. The presence of sagebrush stumps indicates that the historical plant community was rangeland, preventing the misclassification of historic coniferous forests (often more than 100 years old).

Transition T1A

State 1 to 2

The Reference State transitions to the Altered State if mid-statured bunchgrasses, by dry weight, decrease to below 20 percent or if bare ground cover is increased by 15 percent. The driver for this transition is the loss of taller bunchgrasses, which creates open areas in the plant canopy with bare soil. Soil erosion reduces soil fertility, which drives transitions to the Altered State. There are several other key factors signaling the approach of transition T1A: increases in soil physical crusting, decreases in cover of cryptogamic crusts, decreases in soil surface aggregate stability, and/or evidence of erosion including water flow patterns, development of plant pedestals, and litter movement. The trigger for this transition is improper grazing management and/or long-term drought, leading to a decrease in the composition of mid-statured bunchgrasses to less than 20 percent and a reduction in total plant canopy cover.

Transition T1C

State 1 to 3

When mid-stature bunchgrasses are removed from the plant community, the Reference State transitions to the Degraded State, and needle and thread is codominant with short-stature bunchgrasses such as Sandberg bluegrass. This transition differs from T1A in that it is usually quick and associated with disturbances like repeated overgrazing or heavy human traffic. This rapid transition is generally realized where livestock are confined to small pastures for long periods of time, such as horse pastures and calving lots. The driver for this transition is the loss of taller bunchgrasses, which creates openings in the canopy and exposes bare soil. Soil erosion reduces soil health, causing a transition to a Degraded State. There are several other key factors signaling the approach of transition T1C: increases in soil physical crusting, decreases in cover of cryptogamic crusts, decreases in soil surface aggregate stability, and/or evidence of erosion including water flow patterns, development of plant pedestals, and litter movement. The trigger for this transition is improper grazing management, long-term drought, and/or heavy human disturbance.

Transition T1B

State 1 to 4

Healthy plant communities are most resistant to invasion. However, regardless of grazing management, without some form of active weed management (chemical, mechanical, or biological control) and prevention, the Reference State can transition to the Invaded State in the presence of aggressive invasive species such as spotted knapweed, leafy spurge, and cheatgrass. This will occur even if the reference community is thriving. The Central Rocky Mountain Valleys tend to resist invasion by cheatgrass; however, repeated heavy grazing or intense human activities can open the interspaces of the bunchgrass community and allow for encroachment. Long-term stress conditions for native species (e.g., overgrazing, drought, and fire) accelerate this transition. If populations of invasive species reach critical levels, the site transitions to the Invaded State. The trigger for this transition is the presence of aggressive invasive species. The species composition by dry weight of invasive species approaches 10 percent.

Transition T1D

State 1 to 5

The transition from the Reference State to the Conifer Encroached State is driven primarily by long-term fire suppression but heavy grazing may contribute to increased bare ground for seeding sites. Encroachment often occurs fastest within 200 feet of the seed source. The trigger for transition is a conifer stem count higher than 1 per hectare.

Restoration pathway R2A

State 2 to 1

The Altered State has lost soil or vegetation attributes to the point that recovery to the Reference State will require reclamation efforts such as soil rebuilding, intensive mechanical and cultural treatments, and/or revegetation. Low-intensity prescribed fires were used to reduce competitive increaser plants like needle and thread and Sandberg bluegrass. A low-intensity fire will also reduce Wyoming big sagebrush densities. Fire should be carefully planned or avoided in areas prone to annual grass infestation.

Transition T2A

State 2 to 3

As improper grazing management continues, the vigor of bunch grasses will decrease and the shorter grasses and shrubs will increase, contributing to the Degraded State. Prolonged drought will provide a competitive advantage to shrubs, allowing them to become co-dominant with grasses. Shrub canopy will increase. Key transition factors include: an increase in native shrub canopy cover; a reduction in bunchgrass production; a decrease in total plant canopy cover and production; increases in mean bare patch size; increases in soil crusting; decreases in the cover of cryptobiotic crusts; decreases in soil aggregate stability; and/or evidence of erosion, including water flow patterns and litter movement.

Transition T2B

State 2 to 4

Invasive species can occupy the Altered State and drive it to the Invaded State. The Altered State is at risk if invasive seeds and/or other viable material are present. The driver for this transition is more than 10 percent of invasive species.

Transition T2C

State 2 to 5

The transition from the Altered State to the Conifer Encroached State is driven primarily by long-term fire suppression but long-term heavy grazing may contribute to increased bare ground for seeding sites. Encroachment often occurs fastest within 200 feet of the seed source. The trigger for transition is a conifer stem count higher than 1 stem per hectare.

Restoration pathway R3B

State 3 to 1

The Degraded State has lost soil or vegetation attributes to the point that recovery to the Reference State will require reclamation efforts, such as soil rebuilding, intensive mechanical treatments, and/or revegetation. Studies suggest that a mulch with a high carbon to nitrogen ratio, such as wood chips or bark in low moisture scenarios can be beneficial for slow mobilization of plant-available nitrogen (Whitford et al. 1989). Biochar may also be added to the system to improve Soil Organic Carbon (SOC) which should improve Cation Exchange Capacity (CEC), microbial activity, and hydrologic conductivity (Stavi 2012). The drivers for the restoration pathway are the removal of increaser species, restoration of native bunchgrass species, persistent management of invasives and shrubs, and proper grazing management. Without continued control, invasive and shrub species are likely to return (probably rapidly) due to the presence of seeds and/or other viable material in the soil and management-related increases in soil disturbance.

Restoration pathway R3A

State 3 to 2

Since the bunchgrass plant community has been significantly reduced, restoration to the Altered State is unlikely unless a seed source is available. However, if enough grass remains on the site, chemical and/or biological control, in conjunction with proper grazing management, can reduce the amount of shrubs and invasive species and restore the site to the Shortgrass Community. Low-intensity fire can be utilized to reduce Wyoming big sagebrush competition and allow the reestablishment of grass species. Caution must be used when considering fire as a management tool on sites with fire-tolerant shrubs such as rubber rabbitbrush, as these shrubs will sprout after a burn. Broom snakeweed and prairie sagewort may or may not re-sprout depending on conditions (USDA Forest Service, 2011).

Transition T3A

State 3 to 4

Invasive species can occupy the Degraded State and drive it to the Invaded State. The Degraded State is at risk of this transition occurring if invasive seeds or viable material are present. The driver for this transition is the presence of critical population levels (more than 10 percent dry weight of invasive species). The trigger is the presence of seeds or viable material from invasive species. This state has sufficient bare ground that the transition could occur simply due to the presence or introduction of invasive seeds or viable material. This is particularly true of aggressive invasive species such as spotted knapweed. This transition could be assisted by overgrazing (failure to adjust stocking rate to declining forage production), a long-term lack of fire, or an extensive drought.

Transition T3B

State 3 to 5

The transition from the Degraded State to the Conifer Encroached State is driven primarily by long-term heavy grazing and increased bare ground for seeding sites. Encroachment often occurs fastest within 200 feet of the seed

source. The trigger for transition is a conifer stem count higher than one (1) per hectare.

Restoration pathway R4C

State 4 to 1

Restoration of the Invaded State to the Reference State requires substantial energy input. The drivers for the restoration pathway are the removal of invasive species, restoration of native bunchgrass species, persistent management of invasive species, and proper grazing management. Without continued control, invasive species are likely to return (probably rapidly) due to the presence of seeds and/or other viable material in the soil and management-related practices that increase soil disturbance. Sites that have transitioned from the Degraded State to the Invaded State may be severely lacking in soil and vegetative properties that will allow for restoration to the Reference State. Hydrologic function damage may be irreversible, especially with accelerated gully erosion.

Restoration pathway R4B

State 4 to 2

If invasive species are removed before remnant populations of bunchgrass are drastically reduced, the Invaded State can revert to its altered state. The driver for the reclamation pathway is weed management without reseeding. Continued Integrated Pest Management (IPM) will be required as many of the invasive species that can occupy the Invaded State have extended dormant seed life.

Restoration pathway R5A

State 5 to 1

Depending on the level of conifer canopy cover and its impact on rangeland health, restoration efforts may be simply focus on removal of coniferous trees and shrubs to restore the Conifer Encroached State to the Reference State. If following utilizing the phases established by Miller et al management and restoration methods will vary. An large majority of the conifer encroachment in MLRA 44B will fall into the early two phases of Miller's Phases. Phase I may exhibit None-Slight to Moderate departures from rangeland health where removal of the conifers via Brush Management and/or Prescribed fire combined. If mechanical removal of conifers is utilized, no grazing management is needed assuming relatively conservative management had been used prior to treatment. If prescribed fire is utilized, short term grazing deferment and/or rest is suggested. Given a short time removal of a Phase I encroachment will recover to Reference. Proactive pest management is encouraged. Phase II Encroachment may require a more intense mechanical removal of trees/shrubs with prescribed fire not being a feasible method of control as this community may be at risk of catastrophic fire due to canopy density. Phase II displays a Moderate departure from Reference suggesting an overall instability of the site such as reduced herbaceous production, reduced functional/structural groups (e.g. reduced mid-statured bunchgrasses), increase rill frequency and length, and possibly increased bare ground. Increased post treatment grazing management may be necessary. Grazing management may be as simple as short term growing season deferment however long term rest may be necessary in the latter stages of Phase II encroachment. Latter stages of Phase II encroachment will likely require some short term erosion mitigation such as straw wattles as well as range planting and/or critical area planting to re-establish any loss of native herbaceous plants particularly mid-statured cool season bunchgrasses. Phase III Encroachment canopy cover resembles forested sites with larger trees and shrubs. Forest management style tree removal (woody debris and logs removed from the site) will be necessary prior to any prescribed burning as to prevent the fire from burning too hot. The result of a prescribed fire on this site are typically unknown as seed sources of native herbaceous plants are usually limited to small patches. Since the Limy Droughty ecological site for 44B LRU 1 subset C is a dry site, herbaceous plants will likely have been depleted under a Phase III encroachment. This means there is an opportunity for large areas of bare ground, increase rill and in some cases gully erosion. Post treatment will require range planting and/or critical area seeding, erosion control, pest management, and possibly soil carbon amendments (biochar). Grazing management (primarily rest) will be necessary to ensure any new seedling establishment.

Restoration pathway R5B

State 5 to 2

The Conifer Encroached State Phases I and II will generally resemble the Altered State on this site. If following utilizing the phases established by Miller et al management and restoration methods will vary. An large majority of

the conifer encroachment in MLRA 44B will fall into the early two phases of Miller's Phases. Phase I may exhibit None-Slight to Moderate departures from rangeland health where removal of the conifers via Brush Management and/or Prescribed fire combined. If mechanical removal of conifers is utilized, no grazing management is needed assuming relatively conservative management had been used prior to treatment. If prescribed fire is utilized, short term grazing deferment and/or rest is suggested. Given a short time removal of a Phase I encroachment will recover to Reference. Proactive pest management is encouraged. Phase II Encroachment may require a more intense mechanical removal of trees/shrubs with Prescribed Fire not being a feasible method of control as this community may be at risk of catastrophic fire due to canopy density. Phase II displays a Moderate departure from Reference suggesting an overall instability of the site such as reduced herbaceous production, reduced functional/structural groups (e.g. reduced mid-statured bunchgrasses), increase rill frequency and length, and possibly increased bare ground. Increased post treatment grazing management may be necessary. Grazing management may be as simple as short term growing season deferment however long term rest may be necessary in the latter stages of Phase II encroachment. Latter stages of Phase II encroachment will likely require some short term erosion mitigation such as straw wattles as well as range planting and/or critical area planting to re-establish any loss of native herbaceous plants particularly mid-statured cool season bunchgrasses. Phase III Encroachment canopy cover resembles forested sites with larger trees and shrubs. Forest management style tree removal (woody debris and logs removed from the site) will be necessary prior to any prescribed burning as to prevent the fire from burning too hot. The result of a prescribed fire on this site are typically unknown as seed sources of native herbaceous plants are usually limited to small patches. Since the Limy Droughty ecological site for 44B LRU 1 subset C is a dry site, herbaceous plants will likely have been depleted under a Phase III encroachment. This means there is an opportunity for large areas of bare ground, increase rill and in some cases gully erosion. Post treatment will require range planting and/or critical area seeding, erosion control, pest management, and possibly soil carbon amendments (biochar). Grazing management (primarily rest) will be necessary to ensure any new seedling establishment.

Restoration pathway R5C

State 5 to 3

The Conifer Encroached State Phases II and III may resemble the Degraded State on this site. If following utilizing the phases established by Miller et al management and restoration methods will vary. An large majority of the conifer encroachment in MLRA 44B will fall into the early two phases of Miller's Phases. This Restoration Pathway is exceedingly rare as it is typically not cost effect for land managers to manage for a degraded state. Phase I may exhibit None to Slight to a Moderate departures from rangeland health where removal of the conifers via brush management and/or prescribed fire combined. If mechanical removal of conifers is utilized, no grazing management is needed assuming relatively conservative management had been used prior to treatment. If prescribed fire is utilized, short term grazing deferment and/or rest is suggested. Given a short time removal of a Phase I encroachment will recover to Reference. Proactive pest management is encouraged. Phase II Encroachment may require a more intense mechanical removal of trees/shrubs with prescribed fire not being a feasible method of control as this community may be at risk of catastrophic fire due to canopy density. Phase II displays a Moderate departure from Reference suggesting an overall instability of the site such as reduced herbaceous production, reduced functional/structural groups (e.g. reduced mid-statured bunchgrasses), increase rill frequency and length, and possibly increased bare ground. Increased post treatment grazing management may be necessary. Grazing management may be as simple as short term growing season deferment however long term rest may be necessary in the latter stages of Phase II encroachment. Latter stages of Phase II encroachment will likely require some short term erosion mitigation such as straw wattles as well as range planting and/or critical area planting to re-establish any loss of native herbaceous plants particularly mid-statured cool season bunchgrasses. Phase III Encroachment canopy cover resembles forested sites with larger trees and shrubs. Forest management style tree removal (woody debris and logs removed from the site) will be necessary prior to any prescribed burning as to prevent the fire from burning too hot. The result of a prescribed fire on this site are typically unknown as seed sources of native herbaceous plants are usually limited to small patches. Since the Limy Droughty ecological site for 44B LRU 1 subset C is a dry site, herbaceous plants will likely have been depleted under a Phase III encroachment. This means there is an opportunity for large areas of bare ground, increase rill and in some cases gully erosion. Post treatment will require range planting and/or critical area seeding, erosion control, pest management, and possibly soil carbon amendments (biochar). Grazing management (primarily rest) will be necessary to ensure any new seedling establishment.

Additional community tables

Table 9. Community 1.1 plant community composition

Group	Common Name	Symbol	Scientific Name	Annual Production (Lb/Acre)	Foliar Cover (%)
Grass/Grasslike					
1	Mid-Statured Cool Season Bunchgrass			630–1050	
	bluebunch wheatgrass	PSSP6	<i>Pseudoroegneria spicata</i>	395–600	25–55
	needle and thread	HECO26	<i>Hesperostipa comata</i>	60–180	5–15
	green needlegrass	NAVI4	<i>Nassella viridula</i>	60–115	3–6
	rough fescue	FECA4	<i>Festuca campestris</i>	0–100	0–5
	Indian ricegrass	ACHY	<i>Achnatherum hymenoides</i>	0–65	0–2
2	Shortgrasses/Grasslikes			90–210	
	prairie Junegrass	KOMA	<i>Koeleria macrantha</i>	15–60	1–3
	Sandberg bluegrass	POSE	<i>Poa secunda</i>	15–60	1–3
	blue grama	BOGR2	<i>Bouteloua gracilis</i>	10–55	0–3
	Idaho fescue	FEID	<i>Festuca idahoensis</i>	0–30	0–3
	threadleaf sedge	CAFI	<i>Carex filifolia</i>	0–30	0–2
	needleleaf sedge	CADU6	<i>Carex duriuscula</i>	0–15	0–2
3	Rhizomatous Grasses			40–140	
	thickspike wheatgrass	ELLA3	<i>Elymus lanceolatus</i>	40–100	1–3
	western wheatgrass	PASM	<i>Pascopyrum smithii</i>	40–100	1–3
	plains reedgrass	CAMO	<i>Calamagrostis montanensis</i>	0–60	0–1
Forb					
4	Forbs			10–140	
	dotted blazing star	LIPU	<i>Liatris punctata</i>	20–115	0–5
	hairy false goldenaster	HEVI4	<i>Heterotheca villosa</i>	10–60	0–3
	spiny phlox	PHHO	<i>Phlox hoodii</i>	15–60	0–3
	bastard toadflax	COUM	<i>Comandra umbellata</i>	10–60	0–3
	scarlet globemallow	SPCO	<i>Sphaeralcea coccinea</i>	20–60	0–3
	American vetch	VIAM	<i>Vicia americana</i>	10–45	0–3
	fleabane	ERIGE2	<i>Erigeron</i>	20–45	0–2
	desertparsley	LOMAT	<i>Lomatium</i>	0–45	0–2
	common yarrow	ACMI2	<i>Achillea millefolium</i>	0–45	0–2
	rosy pussytoes	ANRO2	<i>Antennaria rosea</i>	0–40	0–2
	Drummond's milkvetch	ASDR3	<i>Astragalus drummondii</i>	0–35	0–2
	locoweed	OXYTR	<i>Oxytropis</i>	0–15	0–1
Shrub/Vine					
5	Shrubs			40–140	
	Wyoming big sagebrush	ARTRW8	<i>Artemisia tridentata</i> ssp. <i>wyomingensis</i>	40–125	5–15
	yellow rabbitbrush	CHVI8	<i>Chrysothamnus viscidiflorus</i>	0–60	0–5
	rubber rabbitbrush	ERNA10	<i>Ericameria nauseosa</i>	0–30	0–3
	spineless horsebrush	TECA2	<i>Tetradymia canescens</i>	0–20	0–1
	plains pricklypear	OPPO	<i>Opuntia polyacantha</i>	0–10	0–1
6	Subshrubs			20–70	
	winterfat	KRLA2	<i>Krascheninnikovia lanata</i>	20–70	1–3

	broom snakeweed	GUSA2	<i>Gutierrezia sarothrae</i>	0–30	0–1
	prairie sagewort	ARFR4	<i>Artemisia frigida</i>	0–20	0–1
	slender buckwheat	ERMI4	<i>Eriogonum microthecum</i>	0–10	0–1

Animal community

The Limy Droughty ecological site of the MLRA 44B provides a variety of wildlife habitat for an array of species. Prior to the settlement of this area, large herds of antelope, elk, and bison roamed. Though the bison have been replaced, mostly with domesticated livestock, elk and antelope still frequently utilize this largely intact landscape.

The relatively high grass component of the Reference Community provides excellent nesting cover for multiple neotropical migratory birds that select for open grasslands, such as the long-billed curlew and McCown's longspur.

Greater sage grouse may be present on sites with suitable habitat, typically requiring a minimum of 15 percent sagebrush canopy cover (Wallestad 1975). The Mid-Staturred Bunchgrass Community is likely to have a minimal sage grouse presence given its low sagebrush canopy cover and forb components. Other communities on the site with sufficient sagebrush cover may harbor sage grouse populations, specifically Community 2.1, where big sagebrush populations are under a reduced fire regime. Also, as sagebrush canopy cover increases under Altered State and Degraded State, pygmy rabbit, Brewer's sparrow, and mule deer use may also increase.

Managed livestock grazing is suitable on this site due to the potential to produce an abundance of high-quality forage. This is often a preferred site for grazing by livestock, and animals tend to congregate in these areas. To maintain the productivity of the Limy Droughty site, grazing on adjacent sites with lower productivity must be carefully managed to ensure that utilization on this site is not excessive. Management objectives should include maintenance or improvement of the native plant community. Careful management of the timing and duration of grazing is important. Shorter grazing periods and adequate deferment during the growing season are recommended for plant maintenance, health, and recovery. Early-season defoliation of bluebunch wheatgrass can result in high mortality and reduced vigor in plants. The opportunity for regrowth is necessary before dormancy to reduce injury.

Grazing season has more influence on winterfat than grazing intensity. Late winter or early spring grazing is detrimental. However, early winter grazing may actually be beneficial (Blaisdell 1984).

Continual non-prescribed grazing of this site will be detrimental, will alter the plant composition and production over time, and will result in the transition to the Altered State. The transition to other states will depend on the duration of poorly managed grazing as well as other circumstances such as weather conditions and fire frequency.

The Altered State is subject to further degradation to the Degraded State or Invaded State. Management should focus on grazing management strategies that will prevent further degradation, such as seasonal grazing deferment or winter grazing where feasible. Communities within this state are still stable and healthy under proper management. Forage quantity and/or quality may be substantially decreased from the Reference State.

Grazing is possible in the Invaded State. Invasive species are generally less palatable than native grasses. Forage production is typically greatly reduced in this state. Due to the aggressive nature of invasive species, sites in the Invaded State face an increased risk of further degradation. Grazing has to be carefully managed to avoid further soil loss and degradation and possible livestock health issues.

Prescribed grazing can be used to manage invasive species. In some instances, carefully targeted grazing (sometimes in combination with other treatments) can reduce or maintain the species composition of invasive species. Grazing may be possible in a degraded state, but it is generally not economically or environmentally sustainable.

Hydrological functions

The hydrologic cycle functions best in the Reference State with good infiltration and deep percolation of rainfall; however, the cycle degrades as the vegetation community declines. Rapid rainfall infiltration, high soil organic matter, good soil structure, and good porosity accompany a high bunchgrass canopy cover of around 80 percent.

High ground cover reduces raindrop impact on the soil surface, which keeps erosion and sedimentation transport low. Water leaving the site will have a minimal sediment load, which allows for high water quality in associated streams. High rates of infiltration will allow water to move below the rooting zone during periods of heavy rainfall. The Mid-Statured Bunchgrass Community should have no rills or gullies present, and drainage ways should be vegetated and stable. Water flow patterns, if present, will be barely observable. Plant pedestals are essentially nonexistent. Plant litter remains in place and is not moved by wind or water.

Improper grazing management results in a community shift to the Mixed Bunchgrass Community. This plant community has a similar canopy cover, but the bare ground will be less than 15 percent. Therefore, the hydrologic cycle is functioning at a level similar to the water cycle in the Mid-Statured Bunchgrass Community. When compared to the Mid-Statured Bunchgrass Community, infiltration rates are slightly reduced and surface runoff is slightly higher.

In the Shortgrass Community, Degraded State, and the Invaded State, canopy and ground cover are greatly reduced compared to the Reference State, which impedes the hydrologic cycle. Infiltration will decrease and runoff will increase due to reduced ground cover, the presence of shallow-rooted species, rainfall splash, soil capping, reduced organic matter, and poor structure. Sparse ground cover and decreased infiltration can combine to increase the frequency and severity of flooding within a watershed. Soil erosion is accelerated, the quality of surface runoff is poor, and sedimentation increases.

Recreational uses

This site provides some limited recreational opportunities for hiking, horseback riding, big game hunting, and upland bird hunting. The forbs have flowers that appeal to photographers. This site provides valuable open space.

Wood products

This site is not suitable for wood products.

Inventory data references

Information presented was derived from the site's Range Site Description (Limy Droughty 15-19 inch P.Z. Northern Rocky Mountain Valleys, South, East of Continental Divide), NRCS clipping data, National Resource Inventory (NRI) data, literature, field observations, and personal contacts with range-trained personnel (i.e., used professional opinion of agency specialists, observations of land managers, and outside scientists).

References

. Fire Effects Information System. <http://www.fs.fed.us/database/feis/>.

. 2021 (Date accessed). USDA PLANTS Database. <http://plants.usda.gov>.

Arno, S.F. and G.E. Gruell. 1982. Fire History at the Forest-Grassland Ecotone in Southwestern Montana. *Journal of Range Management* 36:332–336.

Barrett, H. 2007. *Western Juniper Management: A Field Guide*.

Bestelmeyer, B., J.R. Brown, J.E. Herrick, D.A. Trujillo, and K.M. Havstad. 2004. Land Management in the American Southwest: a state-and-transition approach to ecosystem complexity. *Environmental Management* 34:38–51.

Bestelmeyer, B. and J. Brown. 2005. State-and-Transition Models 101: A Fresh look at vegetation change.

- Blaisdell, J.P. 1958. Seasonal development and yield of native plants on the Upper Snake River Plains and their relation to certain climate factors.
- Blaisdell, J.P. and R.C. Holmgren. 1984. Managing Intermountain Rangelands--Salt-Desert Shrub Ranges. General Tech Report INT-163. USDA Forest Service, Intermountain Forest and Range Experiment Station, Ogden, UT. 52.
- Bunting, S.C., B.M. Kilgore, and C.L. Bushey. 1987. Guidelines for Prescribe burning sagebrush-grass rangelands in the Northern Great Basin. General Technical Report INT-231. USDA Forest Service Intermountain Research Station, Ogden, UT. 33.
- Colberg, T.J. and J.T. Romo. 2003. Clubmoss effects on plant water status and standing crop. *Journal of Range Management* 56:489–495.
- Daubenmire, R. 1970. Steppe vegetation of Washington.
- DiTomaso, J.M. 2000. Invasive weeds in Rangelands: Species, Impacts, and Management. *Weed Science* 48:255–265.
- Dormaar, J.F., B.W. Adams, and W.D. Willms. 1997. Impacts of rotational grazing on mixed prairie soils and vegetation. *Journal of Range Management* 50:647–651.
- Hobbs, J.R. and S.E. Humphries. 1995. An integrated approach to the ecology and management of plant invasions. *Conservation Biology* 9:761–770.
- Kuchler, A.W. 1964. Potential natural vegetation of the conterminous United States.
- Lacey, J.R., C.B. Marlow, and J.R. Lane. 1989. Influence of Spotted knapweed (*Centaurea maculosa*) on surface runoff and sediment yield.. *Weed Technology* 3:627–630.
- Lesica, P. and S.V. Cooper. 1997. Presettlement vegetation of Southern Beaverhead County, MT.
- Manske, L.L. 1980. Habitat, phenology, and growth of selected sandhills range plants.
- Masters, R. and R. Sheley. 2001. Principles and practices for managing rangeland invasive plants. *Journal of Range Management* 38:21–26.
- McCalla, G.R., W.H. Blackburn, and L.B. Merrill. 1984. Effects of Livestock Grazing on Infiltration Rates of the Edwards Plateau of Texas. *Journal of Range Management* 37:265–269.
- McLean, A. and S. Wikeem. 1985. Influence of season and intensity of defoliation on bluebunch wheatgrass survival and vigor in southern British Columbia. *Journal of Range Management* 38:21–26.
- Miller, R.F., T.J. Svejcar, and J.A. Rose. 2000. Impacts of western juniper on plant community composition and structure. *Journal of Range Management* 53:574–585.

- Moulton, G.E. and T.W. Dunlay. 1988. The Journals of the Lewis and Clark Expedition. Pages in University of Nebraska Press.
- Mueggler, W.F. and W.L. Stewart. 1980. Grassland and Shrubland Habitat Types of Western Montana.
- Pelant, M., P. Shaver, D.A. Pyke, and J.E. Herrick. 2005. Interpreting Indicators of Rangeland Health.
- Pellant, M. and L. Reichert. 1984. Management and Rehabilitation of a burned winterfat community in Southwestern Idaho. Proceedings--Symposium on the biology of Atriplex and related Chenopods. 1983 May 2-6; Provo UT General Technical Report INT-172.. USDA Forest Service Intermountain Forest and Range Experiment Station. 281–285.
- Pitt, M.D. and B.M. Wikeem. 1990. Phenological patterns and adaptations in an Artemisia/Agropyron plant community. *Journal of Range Management* 43:350–357.
- Pokorny, M.L., R. Sheley, C.A. Zabinski, R. Engel, T.J. Svejcar, and J.J. Borkowski. 2005. Plant Functional Group Diversity as a Mechanism for Invasion Resistance.
- Ross, R.L., E.P. Murray, and J.G. Haigh. July 1973. Soil and Vegetation of Near-pristine sites in Montana.
- Schoeneberger, P.J. and D.A. Wysocki. 2017. Geomorphic Description System, Version 5.0..
- Smoliak, S., R.L. Ditterlin, J.D. Scheetz, L.K. Holzworth, J.R. Sims, L.E. Wiesner, D.E. Baldrige, and G.L. Tibke. 2006. Montana Interagency Plant Materials Handbook.
- Stavi, I. 2012. The potential use of biochar in reclaiming degraded rangelands. *Journal of Environmental Planning and Management* 55:1–9.
- Stringham, T.K., W.C. Kreuger, and P.L. Shaver. 2003. State and Transition Modeling: an ecological process approach. *Journal of Range Management* 56:106–113.
- Stringham, T.K. and W.C. Krueger. 2001. States, Transitions, and Thresholds: Further refinement fro rangeland applications.
- Sturm, J.J. 1954. A study of a relict area in Northern Montana. University of Wyoming, Laramie 37.
- Thurow, T.L., Blackburn W. H., and L.B. Merrill. 1986. Impacts of Livestock Grazing Systems on Watershed. Page in Rangelands: A Resource Under Siege: Proceedings of the Second International Rangeland Congress.
- Various NRCS Staff. 2013. National Range and Pasture Handbook.
- Walker, L.R. and S.D. Smith. 1997. Impacts of invasive plants on community and ecosystem properties. Pages 69–86 in Assessment and management of plant invasions. Springer, New York, NY.
- Wambolt, C. and G. Payne. 1986. An 18-Year Comparison of Control Methods for Wyoming Big Sagebrush in

West, N.E. 1994. Effects of Fire on Salt-Desert shrub rangelands. Proceedings--Ecology and Management of Annual Rangelands: 1992 May 18-22. Boise ID General Technical Report INT-GTR-313.. USDA Forest Service Intermountain Research Station. 71–74.

Whitford, W.G., E.F. Aldon, D.W. Freckman, Y. Steinberger, and L.W. Parker. 1989. Effects of Organic Amendments on Soil Biota on a Degraded Rangeland. Journal of Range Management 41:56–60.

Wilson, A.M., G.A. Harris, and D.H. Gates. 1966. Cumulative Effects of Clipping on Yield of Bluebunch wheatgrass. Journal of Range Management 19:90–91.

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Approval

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Rangeland health reference sheet

Interpreting Indicators of Rangeland Health is a qualitative assessment protocol used to determine ecosystem condition based on benchmark characteristics described in the Reference Sheet. A suite of 17 (or more) indicators are typically considered in an assessment. The ecological site(s) representative of an assessment location must be known prior to applying the protocol and must be verified based on soils and climate. Current plant community cannot be used to identify the ecological site.

Author(s)/participant(s)	Grant Petersen
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Date	03/01/2020
Approved by	Grant Petersen
Approval date	
Composition (Indicators 10 and 12) based on	Annual Production

Indicators

- Number and extent of rills:** Rills are not present in the reference condition on slopes less than 20 percent. Slopes greater than 20 percent rills may exist but will be extremely rare and less than one (1) foot.

- Presence of water flow patterns:** Water flow patterns are rare in the reference condition. If present, they are most likely to occur on steeper slopes (20 percent) and are inconspicuous, disconnected, and very short in length.

- Number and height of erosional pedestals or terracettes:** Pedestals are not evident in the reference condition.

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4. **Bare ground from Ecological Site Description or other studies (rock, litter, lichen, moss, plant canopy are not bare ground):** Bare ground is seven (7) to 15 percent. Bare ground refers to exposed mineral soil not covered by litter, rock, basal cover, plant cover, standing dead, lichen and/or moss.
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5. **Number of gullies and erosion associated with gullies:** Gullies are not present in the reference condition.
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6. **Extent of wind scoured, blowouts and/or depositional areas:** Wind scoured, or depositional areas are not evident in the reference condition.
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7. **Amount of litter movement (describe size and distance expected to travel):** Movement of fine herbaceous litter may occur within less than a foot from where it originated.
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8. **Soil surface (top few mm) resistance to erosion (stability values are averages - most sites will show a range of values):** Soil Surface Stable with Stability Ratings of 4-6 under canopy and 3-5 under canopy gaps. Biotic crusts and or root mats may be present.
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9. **Soil surface structure and SOM content (include type of structure and A-horizon color and thickness):** Soil structure is weak fine to moderately fine granular. A horizon three (3) to five (5) inches thick, light to dark grey-brown color (Value of 4 or less, Chroma 3 or less) Official Series Description (OSD) for characteristic range.
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10. **Effect of community phase composition (relative proportion of different functional groups) and spatial distribution on infiltration and runoff:** Evenly distributed across the site, bunchgrasses improve infiltration while rhizomatous grass protects the surface from runoff forces. The Limy Droughty ecological site is well drained and has a high infiltration rate. An even distribution of mid stature bunchgrasses (70-75 percent), cool season rhizomatous grasses (5-10 percent), shortgrass (10-15 percent), forbs (1-10 percent), shrubs (5-10 percent), and subshrubs (0-5 percent)
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11. **Presence and thickness of compaction layer (usually none; describe soil profile features which may be mistaken for compaction on this site):** Not Present
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12. **Functional/Structural Groups (list in order of descending dominance by above-ground annual-production or live foliar cover using symbols: >>, >, = to indicate much greater than, greater than, and equal to):**
- Dominant: Dominant: Mid-statured, cool season, perennial bunchgrasses
- Sub-dominant: perennial shortgrasses and grasslikes > rhizomatous grasses > shrubs ≥ forbs > subshrubs
- Other:
- Additional:
-

13. **Amount of plant mortality and decadence (include which functional groups are expected to show mortality or decadence):** Mortality in herbaceous species is not evident. Species with bunch growth forms may have some natural mortality in centers.
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14. **Average percent litter cover (%) and depth (in):** Total litter cover ranges from 20 to 30 percent. Most litter is irregularly distributed on the soil surface and is not at a measurable depth.
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15. **Expected annual annual-production (this is TOTAL above-ground annual-production, not just forage annual-production):** Average annual production is 1200. Low: 900 High 1400 lbs per acre. Production varies based on effective precipitation and natural variability of soil properties for this ecological site.
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16. **Potential invasive (including noxious) species (native and non-native). List species which BOTH characterize degraded states and have the potential to become a dominant or co-dominant species on the ecological site if their future establishment and growth is not actively controlled by management interventions. Species that become dominant for only one to several years (e.g., short-term response to drought or wildfire) are not invasive plants. Note that unlike other indicators, we are describing what is NOT expected in the reference state for the ecological site:** Potential invasive (including noxious) species (native and non-native). Invasive species on this ecological site include (but not limited to) annual bromes., spotted knapweed, yellow toadflax, leafy spurge, ventenata, crested wheatgrass, etc.

Native species such as Rocky Mountain juniper, ponderosa pine, Douglas fir, broom snakeweed, rabbitbrush species., blue grama, Sandberg bluegrass, etc. when their populations are significant enough to affect ecological function, indicate site condition departure.

17. **Perennial plant reproductive capability:** In the reference condition, all plants are vigorous enough for reproduction either by seed or rhizomes in order to balance natural mortality with species recruitment. Density of plants indicates that plants reproduce at level sufficient to fill available resource.
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